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REVIEW





Reviving Contaminated Soils: Microbe-Aided Phytoremediation for Sustainable Metal Pollution Cleanup

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ABSTRACT: Soil metal pollution is a global issue due to its toxic nature affecting ecosystems and human health. This has become a concern since metals are non-biodegradable and toxic. Most of the reclamation methods currently used for soils rely on the use of physical and chemical means, which tend to be very expensive and result in secondary environmental damage. However, microbe-aided phytoremediation is gaining attention as it is an eco-friendly, affordable, and technically advanced method to restore the ecosystem. It is essential to understand the complex interaction between plants and microbes. The primary function of plant growth-promoting bacteria (PGPB) is to stimulate plant development, aid in metal elimination, and reduce their bioavailability in the soil. These microbes regulate phytohormones, stimulate processes such as phytoextraction and phyto-stabilization, and improve the uptake of essential nutrients, such as nitrogen and phosphorus. PGPBs secrete a range of enzymes and chemicals, fix nitrogen, solubilize minerals, increase the bioavailability of nutrients under diverse biological environments with high salinities, excessive metal-contaminated soil, and organic pollutants, increase the soil fertility and help in the reclamation of agriculture and regenerate the native flora. The integration of CRISPR-Cas9 gene-editing technology with microbial-aided phytoremediation and the use of genetically modified microbes with nanomaterials further enhance the efficacy of the approaches in polluted environments for sustainable restoration of the soil.

KEYWORDS: Phytoremediation; heavy metals stress; abiotic stresses; phytoextraction; PGPB

1 Introduction

The accelerated pace of urbanization and industrial growth creates significant environmental concerns, particularly about metal contamination [1]. The elevated concentration of harmful metals in soils raises critical issues for both environmental safety as well as human well-being [2]. Highly beneficial and strongly recommended is the use of chemical and physical remediation methods to address soil metal pollution. The synthetic approaches have the highest frequency of changing the qualities of the soil, have high labor intensiveness, and destroy the soil microorganisms, thereby enhancing the impurities to the ecosystem [3]. Plants-based remediation is an eco-friendly approach to minimize the lethal effects of contaminants in the ecosystem. The strategy entails the application of plants together with the soil microbiota [4]. In contrast to conventional, phytotechnique approaches are recognized as being reasonable and appropriate for extensive use in the field [5,6]. Although a lot of improvements have been observed over the last three decades, the



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approaches are yet under development [7]. The use of bio-agents, sequestering materials, and inoculation of microorganisms, are some of the practices that are considered essential for the effective bio-remediation of soils [8]. To restore ecosystem variability and functionality, microorganisms play a vital role, making it essential to fully comprehend the association of soil microbiome [9,10]. Metal-tolerant plants can thrive in metal-contaminated soil by enhancing their uptake [11].

This article brings out the significance of the remediation of metal-contaminated lands and the crucial function of microbes. Abiotic stresses such as extreme weather or climate change increase soil deterioration rates; the result of which has led to losses of natural ecosystems over the past several centuries [12]. Overuse of fertilizer poses dangerous consequences on soil sustainability. Although a major portion of the global food demand is maintained by land, the solution henceforth must be curative in an attempt to control land degradation. Mankind's priorities have destroyed the environment and ecosystems, lessening the output of farming and balanced living. The main causes of soil degradation include overgrazing, intensive farming, forestry, wood production for fuel, and modernization. Cultivating energy crops in these areas can aid in land restoration and significantly lower greenhouse gas emissions [13]. Much of the Earth's surface is impacted by various abiotic stresses, including contamination with toxic heavy metals (HMs), organic solvents, increased acidity, high salinity, and water scarcity.

Much of the Earth's surface is impacted by numerous abiotic stresses, involving contamination by toxic heavy metals (HMs), organic solvents, high acidity, elevated salinity, and water scarcity [14]. Salinity stress or HM toxicity, pH stress, water scarcity, and other negative factors highly hinder the growth of plants. The interruption of plant physiological activities under stress leads to the loss of leaves, wilting, and less water transpired from leaves. Under such conditions, a drop in turgor pressure limits cell expansion under stress. The loss of turgor pressure is one of the more sensitive biological processes that limit cells from growing under stressful conditions.

Hydrodynamic changes stimulate the enzymatic synthesis of compounds and help to counteract the effects brought about by stress through balancing cellular water [15]. Plants use avoidance and tolerance as strategies to tolerate different stresses, and some species have been reported for their potential to remediate polluted soils [16]. PGPB increases the remediation potential of plants by producing hormones and metabolites, solubilization of minerals, fixation of nitrogen, and protection of plants from infections. PGPB also assists plants in dealing with both biotic and abiotic stresses [17]. Phytoremediation techniques effectively remove pollutants from the environment with less secondary waste, in an economical and eco-friendly way [18]. Such techniques include phyto-stabilization, rhizo-degradation, phyto-desalination, phytovolatilization, phyto-filtration, phytoaccumulation, phyto-transformation, phytoextraction and phytodegradation [19]. While the decomposition of plants and rhizo-degradation deal with biological contaminants, techniques like phytoextraction, phyto-filtration, and phyto-stabilization are mostly applied for soils contaminated with heavy metals. This involves plant species with high efficiency in extracting or immobilizing metals, which possess a strong tolerance to metal contamination, are applied in phytoremediation [20]. The integration of CRISPR-Cas9 gene-editing technology with microbial-aided phytoremediation further enhances the efficacy of the approaches in polluted environments. Genetically modified microbes along with nanomaterials, further accelerate the effectiveness in detoxifying contaminated soils. Moreover, the article brings out the synergistic role of PGPB in phytoremediation and how it could alleviate abiotic stresses and enhance metal uptake, which would be more sustainable and advanced in the restoration of soil.

2 Role of PGPB in Alleviating Abiotic Stresses

Abiotic and biotic stresses decrease the agricultural output. Productivity declines by 50% under abiotic stress and by 30% under biotic stress [21]. It is generally known that PGPBs can reduce the harmful effects of extreme stress on vegetation. In the natural habitat, plants collaborate with a variety of microorganisms, from several kingdoms and domains, comprising of viruses, bacteria, fungi, and archaea. Microorganisms (PGPB) yield valuable compounds, like phytohormones, which shield plants from extreme conditions, whereas the rhizosphere provides the ecological habitats and nutrients for the emergence of microbiota [22]. By using PGPB products, contaminated and unusable land can be turned into fertile ground, which is beneficial for plant growth [23,24].

2.1 PGPB Mitigates Salinity Stress

Translocation of sodium to vesicles lowers the concentration of salts in the cells and is the primary mechanism by which plants tolerate salt [25]. According to research, PGPB phytoremediation is linked to higher expression of the Salt Overly Sensitive 1 (SOS1) and the other genes related to the SOS trail. To increase salt resistance, phytohormones like ethylene, salicylic acid, and abscisic acid are synthesized more often in response to salinity stress. These hormones are liable for stimulating the signaling pathway of several genes [26]. It is reported that inoculation of soil with Bacillus aryabhattai H19-1 and Bacillus mesonae H20-5, PGPB strains may boost the working of antioxidant enzymes, the breakdown of abscisic acid, and the increase of proline under saline stress [27,28]. According to research done on uninoculated plants and plants infected with mutant Pseudomonas species, the manufacture of ACC deaminase (lowers the level of ethylene) by PGPB is the technique that allows plants to survive salinity. It is reported by Girolkar et al. [29], that Streptomyces, Arthrobacter, and Bacillus sp. increased the root growth in wheat, maize, and rice crops. In another study, B. pumilus FAB10 (a salt-resistant phosphate solubilizing bacteria) enhanced the yield of wheat under 25 dSm⁻¹ NaCl stress. Phosphate solubilizing strains enhanced the growth of shoots in pepper and rice under 20 and 15 dSm⁻¹ NaCl [30]. The synthesis of exopolysaccharides (EPS) and the development of biofilms are crucial defensive mechanisms under salt stress. The fresh weight of the chickpea increased by 153% and 177%, respectively, after being inoculated with strain Planococcus rifietoensis RT4 and Halomonas variabilis HT1 of bacteria at a 100 mM NaCl concentration [31].

2.2 Plant Growth Increased by PGPB on High-Salinity Marginal Land

By producing ACC deaminase and biofilms that cover the outside of the roots, which reduce ethylene precursors, PGPB aids plants in alleviating salinity stress in marginal lands [32]. It also increases the effectiveness of water consumption by controlling transpiration, regulating stomatal conductance, and lowering the concentrations of ROS in inoculated species [33]. Inoculating *Pseudomonas* species that produce ACC deaminase in barley and oats for phytoremediation of saline soils, the plants' roots grew by 200% and 50%, respectively, and their shoot biomass increased by 100%–1500% [34]. Also, *Novosphingobium* sp. HRIa and *Pseudomonas putida* KT2440 enhanced *Citrus macrophylla* development. Under salt stress, the strain KT2440 avoided stem chloride and proline accumulation, and the strain HRIa enhanced IAA accumulation in leaves (Fig. 1) [35]. Similarly, increased plant development and decreased salinity were observed in spinach (*Spinacia oleracea* L.) after inoculation with chitinolytic (*Sanguibacter* spp., *Pseudomonas* spp., *Bacillus* spp.) and halotolerant (*Pseudomonas* spp., *Thalassobacillus* spp.) bacterial strains with high antifungal activity. On marginal fields with organic compound contamination, PGPB increased plant growth. The efficacy of phytoextraction is typically lesser than that of phytodegradation or phyto-stimulation mainly because many organic contaminants have strong repellent characteristics [36]. Bacteria increase their ability to biodegrade by using the metabolites that plants release as carbon sources. By doing so, the contaminant's stress is

reduced and plant growth is promoted. High hydrocarbon-resistant bacteria which break down organic pollutants aerobically, are among the PGPB strains beneficial in phyto-stimulation technique. Environmental organic pollutants that can degrade petroleum hydrocarbons are the most common type of pollution in any country [37]. PGPB is commonly obtained from the autochthonic microbiota, for soil bioaugmentation and exogenous pools of microorganisms. Although there is a diversity among these strains, the rhizosphere bacteria are thought to be the best at degrading hydrocarbons [38]. These bacterial strains produce biosurfactants that increase the bioavailability of hydrocarbons, essential for biodegradation. The Bassia scoparia in conjunction with rhizosphere microorganisms exhibited the high efficacy of phyto-stimulation in soil contaminated with crude oil [39]. Cajanus cajan, rhizospheric bacteria, and Zea mays L. were used to phytoremediate petroleum oily sludge, Italian ryegrass was used to break down petroleum hydrocarbons, similar effects of PGPB were seen supported by alkane-degrading bacterial strains [40], and the degradation of diesel contaminants by Zea mays. The cleanup of organic compounds in polluted areas may also use phytostimulation as an additional technique. Using plant-beneficial bacteria, phyto-stimulation helped switch grass to eliminate polychlorinated biphenyls (PCBs). In a study, Burkholderia xenovorans LB400 was used as an adjunct to the key treatment, i.e., phytoextraction [41]. It is feasible to use a range of biotechnological methods to improve phytoremediation, such as changing the genes for HM transporters and the mechanisms that allow them to be absorbed, as well as boosting the production of HM ligands. The HM transporter gene was overexpressed in Arabidopsis thaliana to increase sensitivity and accumulate Pb and Cd (YCF1). Likewise, Nicotiana tabacum NtCBP4 protein overexpression in transgenic plants led to increased Pb accumulation and hypersensitivity. For HM detoxification, ligands that bind to HM, glutathione, phyto-chelatins, and cystinerich peptides like metallothioneins are used. Significant Cu accumulation in roots was seen in peas as a result of the metallothionein (PsMTA) being overexpressed. Similar, results were seen in transgenic Brassica juncea that had the gene for *E. coli* GSH synthetase overexpressed [42].

2.3 Role of PGPB in Drought-Tolerance of Plants

As a result of climate change, water shortage would impact more areas of the world as climate change is predicted to increase by 1.6°C in semi-arid areas of South Africa and 0.2° C– 0.5° C per decade for Asia [43]. Reactive oxygen species (ROS) are accumulated in plants as a result of a signaling cascade activated by drought, and in the absence of detoxifying systems, these ROS can harm proteins, cell membranes, and DNA [44]. In a study, *Azosprillum brasilense* was used to inoculate *Urochloa ruziziensis* leaves, which improved the level of CAT and POD in the plant tissues and enhanced the plant's resistance to water shortage [45]. Exopolysaccharides secreted by PGPB are involved in the manufacture of ACC deaminase, which assists the plant in coping with drought stress. The *Methylobacterium oryzae* (LMG23582(T)) strain stimulated the growth of shoots (32%) and roots (51%), enhanced photosynthetic activity (4.85 μ molCO₂/m²/s), and improved harvest index (4-fold) under drought stress in lentil plants, The stress mitigation effect was due to microbial cytokinins delivered to plants [46].



Figure 1: Methods to stimulate plant growth promotion under abiotic stress by plant growth promoting bacteria (PGPB)

2.4 Role of PGPB in Inducing HM-Tolerance in Plants

According to research by Liu et al. [47], HM (Heavy metal) pollution affects about 20 Mha of land worldwide. HMs are naturally occurring in the Earth's shell and are used excessively in industry. Exposure to HMs can have an impact on key enzymes and cellular organelles in plants. PGPB can improve plant biomass while reducing the adverse properties of HM exposure on plants (Fig. 2) [48] either by aiding phytoextraction or by changing it into a form that isn't bioavailable [49]. The PGPB may develop tolerance to HMs by developing metal-protein complexes or by using methylation, demethylation, and biotransformation techniques [49], thus reducing the lethal effects of Pb, Ni, Cu As, Cd, Zn by preventing their accumulation in the aerial plants of Alnus firma [50]. Maize plants inoculated by Proteus mirabilis strain T2Cr and CrP450 improved Cr tolerance and decreased Cr toxicity [51]. The metal-resistant PGPR *Pseudomonas* sp and Bacillus enhanced the ability of sunflowers to hyperaccumulate metals and caused a 1.7 to 2.5-fold rise in zinc and cadmium in the shoots [52]. The yellow stripe-like (YSL), copper (COPT/Ctr), P1Btype metallic ATPase (HMA), cation efflux, Zn- and Fe-controlled ZIP transporter protein, ZIF1 carrier or Zn-induced facilitator1 families are examples of Zn-related transporters that regulate Zn absorption, interpreting, intracellular transport, and efflux. The ZNT family of micronutrient transporters includes zinc and iron-tolerant transporters like proteins to increase plants' resistance to heavy metals. Similarly, by enhancing nutrient uptake and controlling the absorption of Zn at the gene level, Arbuscular mycorrhiza fungi (AMF) improved E. grandis resistance to elevated stress of Zn. AMF resulted in the upregulation of ZNT:4, COPT/Ctr:2, YSL:3, CE:1 genes and the downregulation of ZNT:9, COPT/Ctr:2, YSL:3, ZIFL:4, CE:1 genes under increased Zn soil environment [53].



Figure 2: Plant growth-promoting bacteria stimulate phytoremediation of metal-polluted soil by accelerating plant growth-enhancing parameters

2.5 PGPB's in Metal Phytoremediation

The effectiveness of phytoremediation is affected by extremely hazardous pollutants or other stressed factors. Using microorganisms either exogenously or endogenously can sustain plant growth in these stressful situations and aid in the mobility of pollutants in soil, phytoremediation can be made successful (Fig. 3) [54]. Mello et al. [55] isolated eleven phosphate-solubilizing bacteria from the quinoa rhizosphere, all of which produced plant growth-promoting substances and showed tolerance to various heavy metals and salinity. Of these, *Bacillus atrophaeus* S8 and *Enterobacter asburiae* QB1 promoted enhanced seed germination in quinoa and increased seedling growth, with great potential to be used as inoculants for enhanced quinoa growth in salty or heavy metal-contaminated soils.

In soils that have been affected by HMs, microbes could promote plant development due to the mobility of HMs or the biological transformation into less dangerous chemical complexes of elements. In a study, endophytic Cd-resistant *P. fluorescens* Sasm05 increased *Sedum alfredi*'s growth and capacity to accumulate Cd by upregulating the gene expression maintaining Cd absorption and transportation by plants, increasing the production and raising IAA levels. According to Ke et al. [56], PGPR inoculated ryegrass showed improved growth and reduced Cu- and Cd uptake. The lead uptake increased (ImM Pb) in *Lathyrus*

sativus plants inoculated with I5 strains, with enhanced chlorophyll retention, carotenoid levels, and the upsurge of antioxidant activity suggesting an improved tolerance to metal uptake [57]. *Klebsiella pneumoniae* HR1 biosorb Cd(II) enhanced *Vigna mungo* plants' tolerance to Cd(II) stress, through systemic tolerance mechanisms [58]. In a study, Zn-tolerant bacterium *Serratia* sp. accumulated very high concentrations of Zn and produced large quantities of EPS (extracellular polymeric substances) helping in the sorption of Zn through complexation. Further, secretion of growth hormones, inhibitory compounds, and solubilization of essential nutrients proved it a good bioenhancer for maize crop cultivation [59].



Figure 3: Phytoremediation of metal contaminated soil by plant growth promoting bacteria (PGPB)

In *M. lupulina*, the chimeric strain of *Sinorhizobium meliloti* enhanced development as well as copper tolerance, with an improved antioxidant defense system [60]. It is reported by Akhtar et al. [61] that *Bacillus* sp. CIK-516 inoculation facilitated radish growth with bacterial-assisted phytoextraction of Ni in contaminated soils. *Streptomyces pactum* (Act12) improved growth metal uptake in roots and shoots of wheat plants, making it a promising eco-friendly solution for phytoremediation in contaminated soils (Table 1) [62]. *Streptomyces rapamycinicus* and *S. cyaneus* promoted sunn hemp growth and Cd accumulation in polluted soils [63]. The *Providencia* sp. bacteria strain 7MM11, produced high levels of Indole-3-acetic acid (IAA), enhancing growth in tomato plants. *Brassica oxyrrhina* accumulated Cu and Zn under HMs stress and drought stress due to IAA synthesis of metal-resistant *P. libanensis* TR1 and *P. reactans* Ph3R3 [64]. In a study inoculating *B. tymphaea–N. tymphaea* with isolated Variovorax strains significantly enhanced

shoot biomass with enhanced uptake of nickel (Ni) in the plants. Inoculation of endophytic plant growthpromoting bacteria improved the growth and stress tolerance of *Noccaea caerulescens* and *Rumex acetosa* in metal-contaminated soils [65] (Fig. 4).

In the case of Pb, the significance of microbes in plant extraction is further reinforced [66], Cd [67], Mn [68] and Cr [69], all provided evidence supporting the involvement of microorganisms in phytoextraction. *R. communis* and *H. annuus* showed significant growth promotion as a result of *psychrobacter* sp. SRS8 soil biological enhancement with nickel-resistant bacteria and their use increased the effectiveness of Ni phytoextraction [70]. With reduced Cd and Mn heavy metal uptake, making it a potential inoculum for phytoremediation of contaminated soils [71].

Plant/Crop	PGPB	Metal extraction/ Accumulation	Effects	PGP trait	References
Lathyrus sativus	Pseudomonas fluorescens, Luteibacter sp., and Variovorax sp.	Pb phytoextraction Increased Pb tolerance and accumulation in plants	An improvement in photosynthetic pigments biosynthesis, membrane stability, proline accumulation, and soluble sugars	IAA, siderophores	[72]
Medicago lupulina	Sinorhizobium meliloti CCNWSX0020	Increased Cu uptake decreasing Cu stress	Increased biomass	IAA production ACC deaminase Siderophores	[73]
Bornmuellera tymphaea Noccaea tymphaea Alyssum murale (intercropping)	Variovorax paradoxus	Increased Ni uptake	Increase in Biomass	IAA production ACC deaminase P solubilization Siderophores	[74]
Vitis vinifera cv. Malbec	Bacillus licheniformis Micrococcus luteus	Decreasing As toxic effects	Increase in plant Biomass	N fixation P solubilization Siderophores	[75]
Brassica nigra	Kocuria species (CRB15)	P solubilization	Root and shoot growth	IAA production	[76]
Leucaena leucocephala	Sinorhizobium Saheli	Decrease in Cd uptake	Root and shoot growth increase in Biomass	IAA production, N fixation, P solubilization	[77]
Perennial ryegrass tall fescue	Bacillus sp. EhS7 Acinetobacter RA1 Bacillus RA2	Decrease in metal Cu, Cd uptake	Increase in plant biomass, decrease in oxidative stress,	IAA production P solubilization	[78]
Raphanus sativus	Bacillus sp. CIK-516	Ni phytoextraction/uptake	Increased biomass chlorophyll and nitrogen contents	IAA, ACCD, and EPS	[79]
Triticum aestivum	Streptomyces pactum Act 12	Cd, Cu and Z n photo- extraction/uptake	Increase in plant biomass Decrease in antioxidant activities and lipid peroxidation	IAA, siderophores, ACCD	[80]
Zea mays	B. cereus TCU11	Cd, Pb, Cu, Ni, photo- extraction/accumulation in plant tissues, and their movement to aerial parts	Under high temperature increase plant weight, photosynthetic pigments, protein contents, Cd, Pb, Cu Zn and Ni accumulation	IAA, siderophores	[81]
Sorghum bicolor	B. cereus TCR17, Providencia rettgeri TCR21, Myroides odoratimimus TCR22	Cr photo- stabilization/accumulation in plants	Under high temperature increase root length, plant biomass, antioxidants Reduced proline and MDA content	IAA, siderophores	[82]

Table 1: Studies involving plant growth promoting bacteria in stress alleviation

(Continued)

Table 1 (continued)

Plant/Crop	PGPB	Metal extraction/ Accumulation	Effects	PGP trait	References
Noccaea caerulescens, Rumex acetosa	Variovorax sp., Micrococcus sp., Microbacterium sp.	Facilitate Zn and Cd translocation/Phyto- stabilization in plants	Enhanced photosynthetic pigments, and soil nutrient cycling	IAA, ACCD, P solubilization, siderophores	[83]
Cicer arietinum	Pseudomonas sp.	Decreased Ni absorption	Increased biomass and reduced metal absorption	IAA, siderophores	[84]
Vigna mungo	Klebsiella pneumonia	Amelioration of Cd toxicity bioaccumulation	Enhanced plant biomass, length, germination and antioxidant activities	IAA, P solubilization, siderophores	[85]
Solanum nigrum	Bacillus sp. QX8 and QX13	Cd and Pb accumulation/photo- extraction by plants	Increased plant biomass, enzymatic activity	IAA, siderophores, ACCD, P solubilization	[86]
Broussonetia papyrifera	B. cereus HM5, B. thuringiensis HM7	Minimize the lethal effect of Mn (phytoextraction)	Increase in plant biomass, root length, surface area, inhibit lipid peroxidation, Decrease MDA content, antioxidant enzyme activity in leaves,	IAA, P solubilization, siderophores	[87]
Zea mays	Serratia sp. ZTB	Decrease Zn phytotoxicity (photo-stabilization), increase in Zn accumulation in host plants	Increase in plant growth	IAA, ACCD, siderophores, and P and K solubilization	[88]
Brassica juncea	S. pactum Act12, B. subtilis, B. licheniformi	accumulation/photo- extraction of Cd and Zn	Promote microbial community, enzymes activity, and plant biomass	P solubilization	[89]
Lolium perenne	Bacillus sp. TZ5	Decreased Cd accumulation/photo- stabilization	Increased plant biomass	IAA, P solubilization	[90]
Chenopodium quinoa willd.	B. atrophaeus GQJK17 S8, E. asburiae QB1	Improved plant tolerance to Cu and Cd	Enhanced the germination rate, seedling biomass and growth vigor index	IAA, siderophores, P solubilization	[91]
Z. mays	Providncia sp.	Decrease Cr translocation/photo- stabilization	Under water stress increase the plant growth, pigments, protein, phenolics and relative water content; Decrease lipid peroxidation, proline, superoxide dismutase activity	IAA, ACCD, siderophores	[92]



Figure 4: Mechanisms involved in bacterial-assisted phytoremediation

By enhancing plant tissues' resistance to Ni by PGPB activity, the microbes further boosted Ni bioaccumulation in them [93]. The bacterial strain *Sinorhizobium saheli* YH1, isolated from V-Ti magnetite tailings, produced IAA, solubilized phosphate, and enhanced plant growth Inoculation of *Bacillus sp.* QX8 and QX13 enhanced *Solanum nigrum* development and increased resistance to Pb and Cd stress [94]. The use of bacterial consortium (*Streptomyces pactum* Act12, *Bacillus subtilis*, and *Bacillus licheniformis*) in mono- and co-applications enhanced the phytoextraction of toxic metals (Cd, Cu, Pb, Zn) using *Brassica juncea* and improved soil enzyme activities, plant growth and antioxidant levels, metal bioavailability and uptake, making it a promising strategy for improving phytoremediation and reducing health risks from contaminated soils. It is reported by Ma et al. [95], Cd-immobilizing PGPB, TZ5, loaded onto biochar for preparation of the biochemical composite material used in bioremediation of Cd-polluted soil. The biochar treatment significantly decreased extractable Cd in soils, increased ryegrass dry weight by 77.78%, decreased Cd concentration in the plant by 48.49%, and increased microbial activity and abundance of the *Bacillus* genus in the rhizosphere, thereby providing a feasible approach for Cd remediation.

The microbes have the possibility of lowering Cr(IV) and reducing As(III) toxicity of soil contaminated with Cr and As [96]. The role of *Providencia* sp. (TCR05) and *Proteus mirabilis* (TCR20) in enhancing *Zea mays* growth and phytoremediation efficiency under drought and chromium (Cr) stress revealed that the bacteria improved plant growth, photosynthesis, and stress tolerance, reduced lipid peroxidation and proline levels, and immobilized Cr in roots, making them promising bio-inoculants for improving plant growth and phytostabilization in Cr-contaminated and drought-prone soils.

In a study PGPR consortium and *M. luteus* reduced arsenic concentration in berries and leaves, while *P. fluorescens* enhanced arsenic tolerance in grapevine [97]. The capacity of plants to transport copper from roots to aerial plant parts has an impact on the HM phytoremediation approach. Plant species with translocation factor values >1 are regarded as strong candidates for phytoextraction, while species with translocation

factor <1 are regarded as suitable phyto-stabilizers [98]. When the concentration of HM in the soil is less and its possible release into other ecosystems does not pose a significant problem, phyto-stabilization is typically used. In a study regarding *Ricinus communis* and *Brassica juncea* infected with Pseudomonas sp. (A3R3) and *Psychrobacter* sp. (SRS8) the phytoremediation technique is employed to phytostabilize Zn. In another study, Sedum plumbizincicola rhizoaccumulation of Cd and Zn was enhanced by the E6S strain like Achromobacter piechaudii [99]. Decreased HM transport and/or accessibility in the ground as well as plants, are the main causes of the microbial increasing phyto-stabilization. Some of these processes include adsorption, bioaccumulation, biosorption, biotransformation, HM complexation and (bio) precipitation, and soil alkalinization. Anionic-charged EPS plays an important part in absorption and cationic-charged metals adsorbed by extracellular membrane functional groups. Metals are then transported to microbial cells via passive or active processes known as biosorption and bioaccumulation, where they are subjected to intracellular precipitation, sequestration, and accumulation. Biotransformation of HMs by microbes can also reduce their bioavailability and mobility. The final step in the process of phytoremediation of fields with HM contamination is called phytovolatilization. HMs that can be bio-transformed into less hazardous volatile molecules, e.g., As, Se, and Hg-contaminated fields, and the effectiveness may also be enhanced by the activity of PGPB. The beneficial bacteria for plants are Agrobacterium sp. and Stenotrophomonas maltophilia were helpful in the efficient phytovolatilization of As in a study using the phytoremediation method on Arundo donax L [100]. For plants to decrease As(V) within their cells to As(III) or subsequently reduce the As into its methylated form or volatilization to reduce toxic As in grains. The exact mechanism underlying this process is dependent on the As-speciation in the soil. Increased soil absorption efficiency is necessary to support microorganisms since it leads to greater As phytovolatilization. In the case of Se contamination, a microbial increase of phytovolatilization was also confirmed. Phytovolatilization depends on the conversion of lethal Se (as selenate) into harmless dimethyl selenide gas. Bacteria boosted Se's high rate of combustion and deposition in plant cells by 35% and 70%, respectively. By switching organomercurials from poisonous and combustible Hg(I) to Hg(II) or from Hg(II) to Hg(0), it is possible to increase the effectiveness of Hgphytovolatilization by microbes as observed by Serratia marcescens BacI56 and Pseudomonas sp. BacI38 inoculation into Zea mays [101].

3 Biotechnological Advancements in Phytoremediation

Recent advancements in genetic engineering and the development of genetically modified microbes (GEMs) have significantly improved bioremediation processes. Genetic modification, especially by CRISPRCas9, has made it easier to manipulate the microbial gene for better heavy metal remediation. Altered metabolic pathways in microbes due to activation of specific genes allow them to focus on specifically target-ing environmental pollution. Research studies on genomics, metagenomics, and other 'omics' technologies have shed more light on the interaction of microbes with pollutants by identification of different genes and pathways important for remediation [102].

The CRISPR-Cas9 gene-editing tool is used to introduce or modify genes within microbes, increasing their ability to degrade metals, such as Cd, Cu, Hg, Ni, and Fe. It has proven effective in engineering microbes to resist metal toxicity and increase the rate of bioremediation. Other gene-editing tools, for instance, TALENs, and ZFNs have also been utilized to design microbes to degrade HMs. TALENs, which stands for Transcription Activator-Like Effector Nucleases, work by using a DNA-binding module that targets specific sequences in the host genome, creating double-stranded breaks with "sticky ends." This process helps ensure genetic modifications are stable. Another similar tool is zinc finger nucleases (ZFNs), which use a 30-amino acid DNA-binding domain to create DSBs at specific locations in the genome. Hybrid nucleases, such as TALENs and ZFNs, have emerged through improvement studies over the years that

scientists have encountered in genetic engineering. These innovations eventually set the stage for CRISPR-Cas technology, a stronger system capable of making multiple gene edits simultaneously with high accuracy. The CRISPR-Cas system operates by using guide RNA, which consists of two types of RNA: crisper-derived RNA (crRNA) and trans-acting antisense RNA (trcRNA). This creates a defense mechanism where the gRNA (guide RNA) guides the Cas9 enzyme to specific DNA sequences, where it creates DSBs by recognizing the matching sequence.

CRISPR has allowed multiple genes in microbes to be targeted at a single time thereby efficiently performing remediation processes using microbes. Genome sequencing, metabolomics, and computational biology have been applied to the identification of potentially useful microbes that could be developed for bioremediation while enhancing their tolerance to pollutants. CRISPR-Cas9 has been successfully used in E. coli and Pseudomonas for targeted remediation of HMs. In parallel, nanomaterials are also being studied for their potential in the enhancement of microbial bioremediation. Nanoparticles, because of their high surface area, reactivity, and ability to interact with heavy metals, reduce and degrade pollutants. Nanomaterials like metal oxide nanoparticles, carbon-based nanomaterials, and nanocomposites enhance electron transfer, which supports microbial activity in breaking down toxic metals. Moreover, Nano biosensors are being developed to monitor the progress of the remediation process. Nanomaterials can also serve as substitutes for conventional biosorbents, with various functional groups improving their efficiency in capturing heavy metals. When combined with bacteria, these nanoparticles enhance the bioremediation process, making it more effective than using microbes alone. The interaction between nanoparticles and microbes depends on several factors, including the properties of the nanoparticles, such as size, shape, and surface coating. The CRISPR-Cas9 gene-editing technology and microbial-aided nanomaterials present a highly effective method to enhance microbial-assisted phytoremediation of contaminated soils, potentially resulting in faster detoxification of heavy metals [103].

4 Future Perspectives

Plants used in bioremediation may effectively utilize contaminated biomass to create a variety of valueadded goods like pigments, chemicals, etc. Burning plant debris can recover metals that can be utilized as starting points for production processes. There are many untapped potential topics for research in the future. The first step is determining how different pollutants interact with one another and how hazardous they are to soil microbial communities when those populations include specific plant species like *Miscanthus* sp. It's also important to look into genetically modified microbes that support phytoremediation, mineral dynamics, carbon from plants, and biodiversity. Due to the extreme sensitivity to a certain metal, several traditional plants could not be used in contaminated soils. However, addressing how climate change affects the dynamics of different chemicals in crops and the discharge of microbial metabolites is urgently needed.

Combining microbes with approaches, such as plants, nanoparticles, or soil additives, further enhances remediation outcomes. Furthermore, pairing microbes with organic or carbon-based materials should also be explored. The behavior of microbes in heavy metal-contaminated environments is essential to understand, and further research is needed to understand the physiochemical, biological, and molecular characteristics that allow microbes to thrive in these conditions. Key research gaps that need to be filled include the understanding of microbial diversity, mechanisms of resistance to metals, and the nature of microbe-soil interaction. The study needs to explore, potential mixed microbial communities, scale-up from laboratory-scale to field application, and the integration with other techniques as well. Addressing the gaps involving safety aspects of the environment, and reduced costs of the technology in targeting heavy metals by phytoremediation will enhance the efficiency and sustainability for large-scale environmental cleanup.

5 Conclusion

Both plants and bacteria aid in the phytoremediation of harmful chemicals. Additionally, bacteria can shield plants from environmental stresses such as HM toxicity, water stress, salt stress, etc., and stimulate growth in plants through a number of PGP processes like hormone and siderophores production, mineral solubilization, nitrogen fixation, and various other mechanisms. By employing microorganisms that are crucial to the phytoremediation of soils, phytoremediation continues to be seen as a relatively easy approach for lowering or bio-transforming contaminants and indirectly increasing plant development. The range of currently employed phytoremediation techniques can be increased by studying bacteria that have changed to be tolerant of high chemicals and their environmental associations, in particular. By employing PGPB and metal-solubilizing bacteria to inoculate the soil, hyperaccumulators' health, biomass, yield, and ability to accumulate metal can all be enhanced. The efficiency of biological microbe-assisted phytoremediation conditions like drought and salinity due to climate change requires more investigation. Additionally, more study is required to comprehend the interactions between microorganisms and plants in an ecosystem that has been contaminated by metal.

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Abbreviations

PGPB	Plant growth promoting bacteria
CRISPR-Cas9	Clustered regularly interspaced short palindromic repeats and CRISPR-associated protein 9
HM	Heavy metal
SOS1	Salt Overly Sensitive 1
EPS	Exopolysaccharides
Pb	Lead
Cd	Cadmium
Cu	Copper
ROS	Reactive oxygen species
Ni	Nickel
As	Arsenic
Zn	Zinc

Cr	Chromium
YSL	Yellow stripe-like
AMF	Arbuscular mycorrhiza fungi
GEMs	Genetically modified microbes
ZFN	Zinc finger nucleases
TALENS	Transcription Activator-Like Effector Nucleases
crRNA	Crisper-derived RNA
gRNA	Guide RNA
DSBs	Double-stranded breaks

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